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Title	Investigating tree foliar preference by the earthworms Aporrectodea longa		
	and Allolobophora chlorotica in reclaimed and loam soil		
Туре	Article		
URL	https://clok.uclan.ac.uk/id/eprint/16181/		
DOI	https://doi.org/10.1016/j.apsoil.2016.10.007		
Date	2017		
Citation	Ashwood, F, Butt, Kevin Richard, Doick, KJ and Vanguelova, EI (2017)		
	Investigating tree foliar preference by the earthworms Aporrectodea longa		
	and Allolobophora chlorotica in reclaimed and loam soil. Applied Soil		
	Ecology, 110. pp. 109-117. ISSN 0929-1393		
Creators	Ashwood, F, Butt, Kevin Richard, Doick, KJ and Vanguelova, El		

It is advisable to refer to the publisher's version if you intend to cite from the work. https://doi.org/10.1016/j.apsoil.2016.10.007

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1 Investigating tree foliar preference by the earthworms *Aporrectodea longa* and

2 Allolobophora chlorotica in reclaimed and loam soil

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15 Abstract

Afforestation can improve the delivery of ecosystem services from reclaimed landfill sites. 16 Tree health is a key determinant of ecosystem service delivery, and is directly impacted by 17 soil quality; which is driven by biological processes in the soil, reliant on leaf litter inputs to 18 function. Different tree species have different litter quality, affecting the degree to which they 19 20 support biological processes in soils and the development of abundant and diverse soil faunal 21 communities. In recognition of their key role in improving soil structure and fertility - key 22 attributes of soil quality, earthworms have often been the subject of research as a part of land reclamation, and these organisms have displayed preferences for specific types of leaf litter. 23 24 This work utilised a choice chamber design to measure the foliar material palatability of two 25 tree species used in land restoration (Alnus cordata and Acer platanoides) as a food source 26 for two common European earthworm species (Aporrectodea longa and Allolobophora chlorotica), and the effect of a reclaimed soil quality on earthworm growth, survival and 27 feeding preferences. The research revealed that both earthworm species initially preferred 28 29 the foliar material of A. cordata over A. platanoides, with the leaves of the latter requiring 30 higher degradation to become palatable to earthworms. The consumption of fresh leaves 31 showed these are a suitable food source for earthworms in choice chamber experiments, which historically have instead relied on senescent leaf litter. Finally, high survival rates of 32 both A. longa and A. chlorotica in the reclaimed soil treatment, in addition to consumption of 33 leaf material of two tree species now widely used on reclaimed landfill sites, demonstrated 34 that these earthworm species are suitable candidates for inoculation to reclaimed land. 35

36 Keywords

37 Italian alder; Norway maple; food preference; landfill; choice chamber; Leaf.

The afforestation of reclaimed land, such as former landfill, can provide improved 39 biodiversity, contribute toward climate change mitigation and adaptation, and improve the 40 delivery of ecosystem services from the site (Bullock et al., 2011). In restored woodland, as 41 with natural woodland, a key source of organic matter addition to the soil is from deposited 42 43 leaf material (Lukac and Godbold, 2011). Tree species differently influence soil quality and soil 44 faunal population development through the quality and quantity of their leaf and root litter 45 (Swift *et al.*, 1979a; Pigott, 1989; Muys *et al.*, 1992; Reich *et al.*, 2005; Rajapaksha *et al.*, 2013). It is therefore of value, when planning woodland restoration, to understand whether the tree 46 47 species planted are likely to provide litter which enables and encourages soil faunal communities to establish, thus supporting soil development and ecosystem service provision 48 49 (Kibblewhite et al., 2008; Rajapaksha et al., 2013). Certain tree species, such as Alnus cordata and Acer platanoides are recommended for planting on reclaimed or industrial land, based 50 on their tolerance for high soil pH and dry soil conditions (Hibberd, 1986). There is, however, 51 currently a paucity of knowledge regarding the interaction between these two non-native 52 53 tree species and native UK soil biota, making these important tree species to investigate 54 further and compare to previous research with similar native species (Rajapaksha et al., 2013). 55

In recognition of their role in improving soil structure and fertility, earthworms have been the subject of research during land reclamation for over 50 years, e.g. (van Rhee, 1969; Curry and Cotton, 1983; Curry, 1988; Butt *et al.*, 1995). Earthworm-mediated mineralisation of organic matter, improvement in nutrient availability, and subsequent improvements in plant growth, are likely to be greater in nutrient-poor soils (Jana *et al.*, 2010). It has been demonstrated that

61 certain earthworm species can distinguish between, and may show a preference for, specific types of leaf litter (Darwin, 1881; Satchell and Lowe, 1967). The chemical composition of litter 62 appears to strongly influence earthworm selectivity, in particular aspects such as the C:N ratio 63 and the content of nitrogen, calcium, lignin and polyphenols (Satchell and Lowe, 1967; 64 Hendriksen, 1990). Earthworm preference has been observed for litter decomposed by micro-65 organisms and fungus, which is more palatable over fresh litter material (Satchell and Lowe, 66 67 1967; Wright, 1972; Cooke and Luxton, 1980; Cooke, 1983). However, there is also an indication that the higher nitrate content in dried green leaves can make these a superior 68 quality food for earthworms than dried, senescent and weathered leaves (Butt, 2011a), yet 69 70 little research has been conducted on this.

71 To date, the majority of laboratory-based earthworm feeding preference studies have either 72 looked at how earthworm species respond to non-tree leaf material, or how the welldocumented earthworm species *L. terrestris* responds to tree litter (Satchell and Lowe, 1967; 73 Doube *et al.*, 1997; Neilson and Boag, 2003). A notable exception is a choice-chamber study 74 by Rajapaksha et al. (2013), which investigated how four European earthworm species 75 76 respond to the litter of a set of common temperate tree species (common alder, common 77 ash, silver birch, sweet chestnut and sycamore) and an exotic Eucalyptus species, using standard Kettering loam soil as a substrate. However, these results do not necessarily 78 represent the activity of the same earthworms in woodland on reclaimed landfill sites, where 79 80 alternative tree species and more inhospitable soil materials are likely to be present. 81 Additionally, and to the authors knowledge, there is currently no information on how a 82 combination of anecic and endogeic earthworm species perform in choice chamber feeding 83 experiments, which would provide results more comparable to field conditions, where these

two ecological groups often coincide (Lavelle, 1983). Whilst senesced leaf litter has been used as an experimental food source in previous choice chamber experiments (e.g. Rajapaksha *et al.*, 2013), due to seasonal unavailability of such material this experiment adopted the use of freshly collected tree foliar material. Since green tree leaves have not been investigated in earthworm choice chambers to date, this provided the opportunity to gather novel information on this material as a food source for earthworms.

90 Therefore, the objectives of this study were to:

- 91 1. Measure the foliar material palatability of two tree species used in land restoration as
- 92 a food source for earthworms, and influence on earthworm mass and survival,
- Measure the effect of reclaimed soil on earthworm mass, survival and foliar selection
 behaviour, compared to a control (Kettering loam) soil,
- 95 3. Obtain data on the above from a combination of endogeic and anecic earthworm96 species relevant to landfill conditions.

98

2. Materials and Methods

99 2.1. Choice chamber and experimental design

This experiment utilised the choice chamber design described by Rajapaksha et al. (2013), 100 101 which is a modified version of Doube et al. (1997) and Rief et al. (2012). This design allows for 102 earthworm food preference to be regularly monitored and quantified by removal of feeding 103 tubes, with minimal disturbance to the central chamber and resident earthworms. The 104 addition of soil to the central chamber rather than moist filter paper (e.g. Doube *et al.*, 1997) provides more natural environmental conditions for endogeic and anecic earthworm species, 105 106 and in this experiment also allowed for comparison between two soil types. This experiment 107 investigated tree foliar preference by two species of earthworm; Allolobophora chlorotica (endogeic) and Aporrectodea longa (anecic); both as monocultures and as a combined species 108 109 treatment. Two soil treatments were investigated; Kettering loam and reclaimed soil. Five 110 trays (replicates) were set up for each combination of soil treatment and earthworm culture (5 replications X 2 soils X 3 earthworm combinations = 30 trays in total). Six food tubes 111 112 containing leaf material from two different tree species litter (e.g. three tubes per tree species) were arranged alternatingly around each tray, and the average mass loss of these 113 per tree species per tray was measured. 114

The choice chamber design consisted of a circular aluminium foil tray (0.16 m diameter and 0.03 m depth) with standard Eppendorf tubes (0.01 m diameter and 0.04 m depth) spaced equally around the choice chamber and embedded into the tray walls as food containers (Figure 1a). To enable the tubes to be affixed to the choice chambers and allow earthworm access to tube contents, the caps were removed from the tubes and a hole of approximately 0.01 m diameter was drilled in each cap (Rajapaksha *et al.*, 2013). An equally-sized hole was then made in the wall of the choice chamber and the caps placed on the inside of the hole,
enabling the tubes to be attached on the outside wall of the choice chamber and held in place
by the caps (Figure 1b). This enables the tubes to be removed from the caps and replaced
without disturbing the contents of the main chamber.

125 [INSERT FIGURE 1]

126 Prior to experimentation, empty Eppendorf tubes were affixed to the choice chambers, and each choice chamber was filled with a soil treatment at 25% moisture content (Figure 2). The 127 two soil treatments were: sterile (heat-treated) Kettering loam topsoil (Boughton Loam, 128 129 Kettering, UK), which is a standard substrate for use in general earthworm experiments and 130 choice chamber experiments (Butt et al., 1994b; Rajapaksha et al., 2013), or sterilised anthropogenic soil materials taken from an afforested reclaimed landfill site in Rainham, UK 131 (Nat. Grid Ref TQ 52572 83192). Sterilised field-collected reclaimed soils were sieved to remove 132 materials >4 mm, then frozen at -5 °C for 7 days to destroy native earthworms and other 133 potential competitors and predators (Butt, 2011b). The average chemical composition of both 134 135 soil treatments at the start of the experiment is provided in Table 1. The reclaimed soil 136 treatment possessed significantly higher pH, conductivity, total C, organic carbon, organic matter (%), C:N ratio and total K (%) than the Kettering loam treatment. The loam soil 137 138 possessed significantly higher total N (%) and Ca, and both soils had similar levels of Na and 139 Mg.

140 [INSERT TABLE 1]

141 [INSERT FIGURE 2]

143 Earthworms were then randomly selected, had masses determined and were allocated to the choice chambers according to the species combination treatments, and sprayed with water. 144 145 This experiment investigated the leaf material preference of two earthworm species: A. longa 146 (anecic) and A. chlorotica (endogeic) with initial individual mean initial masses of 2.30 (SE ± 0.11) and 0.26 (SE ± 0.01) g, respectively. Each earthworm species was introduced to separate 147 choice chambers in the following numbers, according to treatment; monoculture of A. longa 148 149 (4), monoculture of A. chlorotica (20), or a mixed culture of A. longa and A. chlorotica (2 and 10, respectively). These numbers were selected for similar earthworm biomass across choice 150 chambers independent of earthworm treatment, and to ensure a quantifiable rate of leaf 151 152 material removal within the short timeframe of the experiment. All earthworms were collected from agricultural pasture at Walton Hall Farm, Preston, UK (Nat. Grid Ref: SD 55050 153 28100), via digging and hand-sorting of soil. To prevent earthworm escape during the 154 155 experiment, choice chambers were covered with a sheet of aluminium foil held in place by an 156 elastic band. Small holes were made in the foil with a mounted needle to allow for air circulation whilst maintaining soil moisture content. All choice chambers were then stored in 157 158 total darkness in a temperature-controlled incubator at 15°C for a period of 24 hours, to allow earthworms to equilibrate to the experimental conditions. 159

Leaf materials from two tree species were selected for use in this experiment; these were *A*. *platanoides* and *Alnus cordata*. Fresh leaf materials of both species were collected from trees at Ingrebourne Hill Community Woodland (the reclaimed site from which soil materials were obtained). These were separately air-dried and ground using a MAGIMIX 4150W food processor, then sieved to obtain leaf particles of 1 - 2 mm size. Particle size has been shown to influence earthworm selection of food material (Lowe and Butt, 2003), and this size range

166 was chosen to prevent such issues. A sub-sample of both tree species leaf materials was retained for chemical composition analysis. Fresh Eppendorf tubes were individually labelled, 167 168 had masses determined and were filled with dried and sieved leaf particles of either tree 169 species (between 0.2 - 0.3 g per tube), and had mass re-determined. The leaf-filled tubes were then topped-up with water and left to soak for two hours, and inverted on absorbent paper 170 for five minutes to drain excess water. Tubes then had mass re-determined to obtain the wet 171 172 starting mass of the leaf materials. These tubes were then assigned to specific choice chambers and used to replace the empty Eppendorf tubes, thus marking the start of the 173 174 experiment. Three feeding tubes for each species leaf material were placed in alternating 175 positions around each choice chamber, with a total of six tubes per choice chamber (Figure 2). Throughout the experiment, choice chambers were maintained in a temperature-176 controlled incubator at 15°C, in total darkness. 177

178 2.2. Measurements

179 Leaf material removal from feeding tubes was measured every three days, by determining the mass loss (%) of each tube. Earthworm preference was associated with leaf removal. 180 Following mass recording, each tube was then re-attached in the same location. During 181 182 measurement periods, each choice chamber had its foil lid removed and was inspected for signs of dead earthworms, with any mortalities recorded and the remains removed. Soil 183 moisture content was maintained in each choice chamber by spraying each with an equal 184 amount of water during inspection. The experiment was terminated after 27 days, or earlier 185 for any choice chamber when all leaf material had been removed from the feeding tubes. At 186 termination of the experiment, earthworm survival and final masses were recorded for each 187 choice chamber. 188

189 2.3. Statistical Analysis

190 Statistical analysis was performed using the freeware statistical software R, version 3.2.2. 191 "Fire Safety" and the R Studio desktop software, version 0.99.486 (R Core Team, 2015; 192 RStudio Team, 2015). Data were first tested for normality using the Shapiro-Wilk test, which is suited to smaller sample sizes (in this case n=5). All leaf removal data for each species and 193 soil treatment had a normal distribution. To identify feeding preference midway through the 194 experiment, Paired Student's t-test was applied to the leaf removal data at the point at 195 which 50% total leaf material was removed from choice chambers in each tray, as per 196 197 Doube et al. (1997) and Rajapaksha et al. (2013). Two-way repeated measures ANOVA was applied to the complete dataset across all time points, to investigate the influence of 198 experiment duration alongside treatments on earthworm leaf material removal. 199

200 **3. Results**

The choice chambers enabled accurate monitoring of earthworm feeding behaviour, with clear visual and gravimetric evidence of leaf foliar material removal throughout the experiment, and a generally similar pattern for all species combinations across soil treatments.

Table 2 shows earthworm performance across treatments at the start and at termination of the experiment. After 27 days, 100% survival was recorded for *A. longa* across all treatments. *A. chlorotica* had 98-99 % survival in reclaimed soil, but survival was much lower (35-46%) in the loam treatment. *A. chlorotica* lost mass across all treatments (range of -4.0 to -41.0%), *A. longa* lost mass in the monoculture loam treatment combination (-1.89% loss) and gained mass across all other treatment/species combinations (+15.5 to +20.0% gain).

211 [INSERT TABLE 2]

Figure 3 illustrates the pattern of leaf litter removal from choice chambers by all three 212 earthworm species combinations supplied with A. cordata and A. platanoides foliar material 213 214 over 27 days for both soil treatments. All three species combinations showed a clear initial preference for A. cordata leaf material over that of A. platanoides. After 12 days, the rate of 215 A. platanoides leaf material removal by all earthworm species rapidly increased under both 216 soil treatments. Despite the large difference in survivorship between A. chlorotica in the loam 217 and reclaimed soils (35 and 99%, respectively, Table 2), there was little difference in final litter 218 219 removal between treatments (see also Figure 3). Foliar material removal by A. chlorotica monoculture was linear throughout the experiment, although far reduced compared with A. 220 221 *longa* monoculture and the mixed species treatment.

222 [INSERT FIGURE 3]

223 For A. longa monoculture in the loam soil treatment, at 15 days (the point of 50% total leaf 224 removal) the amount of *A. cordata* was significantly less than *A. platanoides* (ANOVA, F (1, 8) 225 = 25.66, p < 0.001, see Table 3). In the reclaimed soil treatment, A. longa displayed a similar 226 pattern of litter removal, which was also statistically significant (ANOVA, F (1, 8) = 9.77, p = 227 0.014). There was also a significant effect of soil on leaf material removal (two-way repeat 228 measures ANOVA, F (1, 16) = 6.39, p = 0.022). The combined species treatment showed a 229 similar, although less pronounced leaf preference result to A. longa monocultures and results were not statistically significant. A. chlorotica showed a clear trend of litter removal, although 230 50% was not reached at termination of the experiment after 27 days. As with the other 231 232 earthworm species treatments, A. chlorotica consumed more A. cordata than A. platanoides 233 leaf material, in both soil treatments.

Table 3 displays the remaining leaf litter (%) at 50% of total litter removal for each series of choice chambers in the experiment, the point of which varied with earthworm species combinations, but did not vary across soil treatments; *A. longa* (15 days) *A. chlorotica* (50% not removed by experiment termination at 27 days), and mixed species (21 days). At the point of 50% removal, *A. longa* monocultures and the mixed earthworm species treatment showed a clear preference for *A. cordata* over *A. platanoides*.

240 [INSERT TABLE 3]

The results of chemical analysis of leaf material at the start and end of the experiment (bulked material remaining in tubes after 27 days, n = 1) are given in Table 4. Both tree species leaf material showed an increase in total N, P, Ca and Mg (%) at termination of the experiment, and a reduction in C:N ratio and total K (%). At the outset, *A. cordata* leaf material had higher total N (%) and lower C:N ratio and Ca (%) than *A. platanoides* leaves.

246 [INSERT TABLE 4]

Using the results for the loam control soil treatment presented in Table 3, the leaf foliar removal data of *A. longa* and *A. chlorotica* can be compared to the litter preference data for these earthworm species presented by Rajapaksha *et al.* (2013). Table 5 shows earthworm preference for *A. cordata* and *A. platanoides* compared with the leaf litter preference list of Rajapaksha *et al.* (2013).

252 [INSERT TABLE 5]

254

4 **4.** Discussion

255 4.1. Earthworm combinations

All three earthworm treatments demonstrated a preference for the foliar material of A. 256 cordata over that of A. platanoides. The anecic species A. longa displayed rapid removal of 257 foliar material - in monoculture this species removed an average of 4.1 mg leaf material/g 258 fresh weight of earthworm/day, compared to 1 mg leaf material/g fresh weight of 259 earthworm/day displayed by A. chlorotica in monoculture. Little data exists in the literature 260 261 regarding an average OM consumption rate for these two earthworm species – however an accepted average range of 12-17 mg grass litter/g fresh mass of earthworm/day has been 262 263 reported for six temperate grassland earthworm species (van Rhee, 1963; Curry and Schmidt, 264 2007). In woodland habitats A. longa feeds directly on leaf litter material on the soil surface, pulling the material into vertical burrows in the soil (Satchell, 1983). By comparison, the 265 endogeic earthworm species A. chlorotica, which primarily feeds on organic matter within the 266 soil, demonstrated a much slower removal of leaf material; yet this species also showed a 267 preference at the outset of the experiment for A. cordata over A. platanoides foliar material. 268 269 Similar trends in relative rates of litter removal from choice chambers was observed by 270 Rajapaksha et al. (2013) for different earthworm species representatives of the same two ecological groupings: L. terrestris (anecic) and A. caliginosa (endogeic). This was attributed to 271 the different feeding behaviours and the differences in physical size between the two species. 272 273 The large difference in survivorship between A. chlorotica in the loam and reclaimed soils 274 resulted in little difference in final litter removal between treatments, indicating that minimal feeding was taking place by the surviving *A. chlorotica* in both soil treatments. This suggests 275 276 that the food quality or type provided is not particularly suited to this species and/or 277 ecological group. This is likely due to the geophagous nature of this species, and as such,

future feeding experiments involving endogeic geophagous species should take this into consideration. As also found by Rajapaksha *et al.* (2013), the current choice chamber design was better suited to larger, litter-feeding earthworm species than smaller, soil-feeding earthworms.

Earthworm body size and food particle size may have also influenced leaf foliar material 282 283 removal. Neilson and Boag (2003) observed a low removal of food by A. chlorotica during a 284 choice experiment, and found that for the six earthworm species investigated, the mass of 285 food removed was positively correlated with earthworm body size. Food particle size has also been demonstrated to influence intake by earthworms, with reduced particle size generally 286 being of greater benefit to smaller earthworms; however the effects of food size on growth 287 and reproduction may be both species and life-stage specific (Boyle, 1990; Lowe and Butt, 288 2003). 289

290 The addition of an anecic earthworm species might be expected to provide benefits to an endogeic earthworm species, through comminution and incorporation of leaf litter into the 291 292 soil where it can be more easily consumed (e.g. Lowe and Butt, 2003). In controlled laboratory 293 experiments, Butt (1998) and Lowe and Butt (1999) investigated the influence of inter- and 294 intra-specific interactions on earthworm growth rates and reproductive output. Results 295 indicated that earthworm mass was generally negatively affected by the presence of other species, however the severity of the negative influence was related to the extent of niche 296 overlap between the species (Lowe and Butt, 1999). They suggested that the greatest 297 298 competitive interaction effects were present between species representing the same 299 ecological group; findings which support observations by Edwards and Lofty (1978) of negative correlations between ecological grouping and the field densities of four UK 300

earthworm species. Lowe and Butt (2002a) found that inter- and intra-specific interactions negatively influenced earthworm growth, maturation and fecundity; and this was again directly related to the extent of niche overlap between pairings. A notable exception was found for *A. chlorotica*, which exhibited enhanced growth and cocoon production in the presence of *A. longa*. It was concluded that the results of earthworm species interactions cannot be predicted simply based on ecological groupings (Lowe and Butt, 2002a).

307 In this experiment, A. longa demonstrated greater increase in final mass when in combination 308 with A. chlorotica, compared with A. longa monoculture, across both soil types. This supports 309 the findings of Lowe and Butt (2002a), whereby mature anecic *L. terrestris* exhibited greatest masses when paired with endogeic earthworm species. However, the mechanism by which 310 311 endogeic earthworms might have a positive influence on anecic earthworm mass is difficult 312 to identify. It may be the case that the greater A. longa final mass change is the result of reduced intra-specific competition between the two species of different ecological groupings 313 for the limited food resources of the close experimental environment (Lowe and Butt, 1999). 314

315 The lack of any clear change in A. chlorotica mass between combined species and 316 monoculture suggests that A. longa did not provide a positive inter-specific relationship to A. chlorotica, e.g. by facilitating A. chlorotica feeding. Lowe and Butt (2002a) identified that 317 juveniles of one ecological group may have a "niche overlap" and subsequent negative 318 interaction with members of another ecological grouping. However, the earthworms used 319 here were all adults, and as such this cannot explain the lack of inter-specific interaction 320 321 observed on A. chlorotica. Interestingly, the mixed earthworm species treatment was almost 322 as effective as the A. longa monoculture at consuming leaf litter. This would seem to suggest that A. chlorotica acted in leaf removal alongside A. longa, however this does not appear to 323

be reflected in earthworm mass data for this species. The results of earthworm mass and leaf
removal rate for the combined earthworm treatment suggest that these species can co-exist
as an inoculum, and therefore represent an appropriate species combination for inoculation
into field experiments on reclaimed landfill.

328 4.2. Leaf palatability

329 The initial preference for A. cordata foliar material over that of A. platanoides indicates 330 greater quality and palatability of this tree species leaf material to the earthworm species in 331 the experiment, particularly A. longa. Previous studies have helped to identify the chemical and physical parameters of litter which influence litter palatability to earthworms. The 332 333 chemical composition of litter appears to strongly influence earthworm selectivity, in particular aspects such as the C:N ratio and the content of nitrogen, calcium, lignin and 334 335 polyphenols (Satchell and Lowe, 1967; Hendriksen, 1990; Reich et al., 2005; Rajapaksha et al., 2013). Generally, higher N and Ca content and a lower C:N ratio have been associated with 336 increased palatability of leaf litter to earthworms (Reich et al., 2005; Rajapaksha et al., 2013). 337 Current results generally fit this trend; at the start of the experiment, A. cordata foliar 338 339 material had higher total N (%) and lower C:N ratio and Ca (%) content than that of A. 340 platanoides. In a similar study, Rajapaksha et al. (2013) found that leaf litter from the least preferred tree species, sweet chestnut (Castanea sativa), demonstrated particularly low 341 levels of nitrogen and calcium, and highest C:N ratio of all tree species investigated: alder (A. 342 343 glutinosa), common ash (F. excelsior), silver birch (Betula pendula), sweet chestnut (Castanea sativa), sycamore (Acer pseudoplatanus), and an exotic Eucalyptus species (Eucalyptus 344 345 nitens). However, the preferred tree species A. cordata had lower calcium content than the 346 less-preferred A. platanoides, which suggests that calcium content may be less important for

347 leaf palatability, compared to other parameters such as N or C:N ratio. Other factors may affect leaf palatability to earthworms besides those already discussed, such as lignin and 348 tannin content (Hendriksen, 1990). Whilst these were not analysed in the present study, 349 literature indicates that A. cordata and A. platanoides foliar material possess a lignin content 350 of 14.9 (SE ± 1.8) % and 10.2 (SE ± 0.3) %, respectively (Dromenach et al., 1994; Hejcmanová 351 et al., 2014). Using these figures, it appears that foliar lignin content was unlikely to explain 352 353 palatability to the earthworms in this study, as also found by Hendriksen (1990) for tree litter 354 palatability to detritivorous earthworms. Hobbie et al. (2014) found the leaf litter of A. platanoides possesses a cellulose and hemicellulose content of 17.8% (± 0.3) and 16.5% (± 355 0.2), however no data could be found in the literature for these variables on A. cordata foliar 356 or litter material, for comparison. It is strongly recommended that these are assessed in 357 future feeding preference studies, as increased cellulose content has been associated with 358 359 higher C:N ratio and therefore a reduction in leaf palatability to earthworms; with a need for 360 a period of weathering to overcome this (Dickinson, 2012).

361 It has been suggested that litter selection by earthworms can be affected by the state of leaf 362 litter decomposition or weathering (Satchell and Lowe, 1967; Hendriksen, 1990). Earthworms 363 have been shown to prefer decomposed litter by fungal and bacterial colonisation over fresh litter (Satchell and Lowe, 1967; Wright, 1972; Cooke and Luxton, 1980; Cooke, 1983; 364 Hendriksen, 1990). Over the course of this experiment, microbial activity may have affected 365 366 leaf foliar chemical composition and palatability to earthworms. Both tree species leaf litter 367 showed increase in total N, Ca and Mg (%) at termination of the experiment, and a reduction 368 in C:N ratio and K (%). Microbial colonisation of decaying leaf litter leading to litter 369 decomposition has been positively related to increase in N concentration and negatively

correlated with C:N ratio, K and lignin concentrations (Swift *et al.*, 1979b; Hendriksen, 1990).
This represents a positive change in the key chemical parameters which are thought to effect
leaf palatability, and likely explains the sudden increase in A. platanoides foliar material
removal by all earthworm treatments mid-way through the experiment (since there was still *A. cordata* leaf material available at this point, the increased consumption of *A. platanoides*material was unlikely due to lack of other food resources).

The use of green leaf foliar material was shown to successfully support earthworm growth and survival, particularly so for the anecic earthworm species *A. longa*. This supports the findings of Butt (2011a), who used dried green *Betula pendula* leaves as feedstock for *L. terrestris* and found that switching from dried senesced leaves to green leaves during a long term experiment resulted in increased *L. terrestris* mass and significantly increased cocoon production. This was attributed to the larger nitrate content in green leaves enabling more rapid protein synthesis for growth and reproduction.

383 4.3. Soil treatments

384 Soil type did not appear to influence earthworm leaf species preference, with the same trend 385 of leaf selection observed for both soil types and earthworm species. There was, however, a slower rate of leaf consumption observed in the reclaimed soil treatment for all earthworm 386 387 species combinations treatments. This may be linked to higher soil organic matter content in the reclaimed soil (5.9%) compared with the loam (4.7%), which may have enabled increased 388 geophagous feeding rather than direct leaf removal in both A. chlorotica and A. longa (Lowe 389 390 and Butt, 2002b). Typically, soil materials on newly reclaimed landfill sites are unlikely to have 391 high levels of organic matter content (Bending et al., 1999). The levels observed in the reclaimed soils used in this experiment may represent the accidental inclusion of root and 392

other dead plant material (and therefore greater levels of labile carbon for earthworm
 utilisation), since the soil was collected from a re-vegetated 10-year-old reclaimed landfill site.

395 In this experiment, A. longa demonstrated 100% survivorship in both soil treatments, whilst A. chlorotica showed higher survival in reclaimed soil compared to Loam (98% and 35% 396 respectively). Both earthworm species displayed tolerance for soil pH of >8.0, which is above 397 398 that typically recommended for these species, and higher than previous research suggest A. longa may tolerate (Baker and Whitby, 2003; Lowe and Butt, 2005). Overall, both earthworm 399 400 species demonstrated high tolerance of the reclaimed soil used in this experiment, supporting 401 the findings of Butt et al. (2004) who recorded sustainable populations of A. longa and A. chlorotica over a period of ten years following inoculation into reclaimed landfill. 402

403 In the Kettering loam treatment, A. chlorotica showed low survivorship and a decrease in final 404 individual mass. This was surprising, since this soil material has been widely successfully used and is recommended as a standard soil for earthworm-focussed laboratory experiments (Butt 405 et al., 1994b; Lowe and Butt, 2005; Rajapaksha et al., 2013). Earthworm survival and activity 406 407 is greatly influenced by abiotic factors, in particular soil temperature and moisture content; 408 however in this experiment these were maintained at optimal levels and are therefore 409 unlikely to explain the A. chlorotica mortality observed (Lowe and Butt, 2005). Starvation of 410 this geophagous species is unlikely to be the cause of death, since the soil organic matter content of the loam used in this experiment (4.7%) was only marginally lower than that used 411 in other experiments (5%) (Butt et al., 1994a; Rajapaksha et al., 2013). It may be the case that 412 413 the loam soils used in this experiment had become contaminated in some manner during storage prior to the experiment. One proposed explanation for the high rate of A. chlorotica 414 mortality is a negative influence of the decomposition of any early mortalities (e.g. from 415

416 stress/adverse soil conditions upon transport to trays) upon the survival of surrounding 417 earthworms in a closed microcosm. There is currently no discussion of this potentially 418 antagonistic effect in the literature, likely due to the difficulty in distinguishing this from other 419 negative environmental conditions triggering earthworm mortality.

420 **5.** Conclusions

The choice chamber experiment described in this study clearly demonstrated that green leaf 421 material is a suitable food source for the earthworm species investigated. Different tree leaf 422 quality impacts on litter palatability to earthworms with A. cordata foliar material of better 423 424 quality than that of A. platanoides, which needed more time to undergo some degradation 425 before it became palatable to earthworms. The earthworm species A. longa and A. chlorotica demonstrated tolerance (survival and mass increase) of the reclaimed soil used in this 426 experiment, as well as a moderate consumption rate (in the case of A. longa) of the leaf 427 material of trees common to reclaimed landfill sites. As such these earthworm species 428 represent suitable candidates for inoculation to reclaimed landfills, where suitable conditions 429 430 prevail.

431 Acknowledgements

The authors would like to thank the Forestry Commission England's Thames Beat team for
providing funding and resources to support this research. We would also like to thank Jack
Forster for helpful advice on statistical analysis.

435 References

Baker, G. H. and Whitby, W. (2003). Soil pH preferences and the influences of soil type and
temperature on the survival and growth of Aporrectodea longa (Lumbricidae). *Pedobiologia*,
47 (5–6), p.745–753.

- Bending, N. A. D., McRae, S. G. and Moffat, A. J. (1999). *Soil-forming Materials: Their Use in Land Reclamation*. London: DETR.
- Boyle, K. E. (1990). The ecology of earthworms (Lumbricidae) in grassland on reclaimed
- 442 cutover peatland and their impact on soil physical properties and grass yield. Unpublished
- 443 Ph. D. thesis, National University of Ireland. Cited in Edwards, CA (ed) Earthworm Ecology. St.
- 444 *Lucie Press, Boca Raton*, JOUR, p.37–64.
- Bullock, J. M., Aronson, J., Newton, A. C., Pywell, R. F. and Rey-Benayas, J. M. (2011).
- 446 Restoration of ecosystem services and biodiversity: Conflicts and opportunities. *Trends in*447 *Ecology & Evolution*, 26 (10), article, p.541–549.
- Butt, K. R. (1998). Interactions between selected earthworm species: A preliminary,
 laboratory-based study. *Applied Soil Ecology*, 9 (1–3), p.75–79.
- Butt, K. R. (2011a). Food quality affects production of Lumbricus terrestris (L.) under
 controlled environmental conditions. *Soil Biology and Biochemistry*, 43 (10), p.2169–2175.
- 452 Butt, K. R. (2011b). The earthworm inoculation unit technique: development and use in soil
- 453 improvement over two decades. In: Karaca, A. (ed.), *Biology of Earthworms*, Berlin: Springer454 Verlag, p.87–105.
- 455 Butt, K. R., Frederickson, J. and Morris, R. M. (1994a). Effect of earthworm density on the
- 456 growth and reproduction of Lumbricus terrestris L. (Oligochaeta: Lumbricidae) in culture.
- 457 *Pedobiologia*, 38, article, p.254–261.
- Butt, K. R., Frederickson, J. and Morris, R. M. (1994b). The life cycle of the earthworm
 Lumbricus terrestris L. (Oligochaeta: Lumbricidae) in laboratory culture. *European Journal of Soil Biology*, 30 (2), article, p.49–54.
- Butt, K. R., Frederickson, J. and Morris, R. M. (1995). An earthworm cultivation and soil
 inoculation technique for land restoration. *Ecological Engineering*, 4 (1), article, p.1–9.
- Butt, K. R., Lowe, C. N., Frederickson, J. and Moffat, A. J. (2004). The development of
 sustainable earthworm populations at Calvert landfill site, UK. *Land Degradation and Development*, 15, article, p.27–36.
- 466 Cooke, A. and Luxton, M. (1980). Effect of microbes on food selection by Lumbricus
 467 terrestris. *Revue d'Ecologie et de Biologie du Sol*, 17 (3), JOUR, p.365–370.
- 468 Cooke, J. (1983). The effects of fungi on food selection by Lumbricus terrestris L. In: Satchell,
 469 J. E. (ed.), *Earthworm Ecology*, London: Chapman & Hall, p.365–373.
- 470 Curry, J. P. (1988). The ecology of earthworms in reclaimed soils and their influence on soil
 471 fertility. In: *Ecology of Earthworms*, p.151–161.
- 472 Curry, J. P. and Cotton, D. C. F. (1983). Earthworms and land reclamation. In: Satchell, J. E.
- 473 (ed.), *Earthworm Ecology*, London: Chapman & Hall, p.215–228.
- 474 Curry, J. P. and Schmidt, O. (2007). The feeding ecology of earthworms A review.
- 475 *Pedobiologia*, 50, article, p.463–477.
- 476 Darwin, C. (1881). *The Formation of Vegetable Mould, Through the Action of Worms, with*477 *Observations on their Habits*. London: John Murray.
- 478 Dickinson, C. H. (2012). *Biology of plant litter decomposition*. Elsevier.
- Doube, B. M., Schmidt, O., Killham, K. and Correll, R. (1997). Influence of mineral soil on the

- palatability of organic matter for lumbricid earthworms: a simple food preference study. *Soil Biology and Biochemistry*, 29 (3/4), p.569–575.
- 482 Dromenach, A.-M., Moiroud, A. and Jocteur-Monrozier, L. (1994). Leaf Carbon and Nitrogen
 483 Constituents of some Actinorhizal Tree Species. *Soil Biol. Biochem*, 26 (5), p.649–653.

Edwards, C. A. and Lofty, J. R. (1978). The influence of arthropods and earthworms upon root growth of direct drilled cereals. *Journal of Applied Ecology*, 15 (3), article, p.789–795.

486 Hejcmanová, P., Stejskalová, M. and Hejcman, M. (2014). Forage quality of leaf-fodder from

the main broad-leaved woody species and its possible consequences for the Holocene

development of forest vegetation in Central Europe. *Vegetation History and Archaeobotany*,
23 (5), p.607–613.

- Hendriksen, N. B. (1990). Leaf litter selection by detritivore and geophagous earthworms. *Biology and Fertility of Soils*, 10 (1), p.17–21.
- Hibberd, B. G. (1986). Forestry Commission Handbook 5: Urban Forestry Practice. London:
 HMSO.

Hobbie, S. E., Baker, L. A., Buyarski, C., Nidzgorski, D. and Finlay, J. C. (2014). *Decomposition*of tree leaf litter on pavement : implications for urban water quality. p.369–385.

- Jana, U., Barot, S., Blouin, M., Lavelle, P., Laffray, D. and Repellin, A. (2010). Earthworms
- influence the production of above-and belowground biomass and the expression of genes
 involved in cell proliferation and stress responses in Arabidopsis thaliana. *Soil Biology and*
- 499 *Biochemistry*, 42 (2), article, p.244–252.
- 500 Kibblewhite, M. G., Ritz, K. and Swift, M. J. (2008). Soil health in agricultural systems.
- Philosophical transactions of the Royal Society of London. Series B, Biological sciences, 363,
 p.685–701.
- Lavelle, P. (1983). The structure of earthworm communities. In: *Earthworm ecology*, Springer, p.449–466.
- Lowe, C. N. and Butt, K. R. (1999). Interspecific interactions between earthworms: A laboratory-based investigation. *Pedobiologia*, 43, article, p.808–817.
- Lowe, C. N. and Butt, K. R. (2002a). Growth of hatchling earthworms in the presence of
 adults: interactions in laboratory culture. *Biology and Fertility of Soils*, 35 (3), article, p.204–
 209.
- Lowe, C. N. and Butt, K. R. (2002b). Influence of organic matter on earthworm production
- and behaviour: a laboratory-based approach with applications for soil restoration. *European* Journal of Soil Biology, 38 (2), article, p.173–176.
- 513 Lowe, C. N. and Butt, K. R. (2003). Influence of food particle size on inter- and intra-specific
- interactions of Allolobophora chlorotica (Savigny) and Lumbricus terrestris. *Pedobiologia*, 47
 (5–6), p.574–577.
- Lowe, C. N. and Butt, K. R. (2005). Culture techniques for soil dwelling earthworms: A review. *Pedobiologia*, 49 (5), article, p.401–413.
- Lukac, M. and Godbold, D. L. (2011). *Soil ecology in northern forests: a belowground view of a changing world*. Cambridge: Cambridge University Press.
- 520 Muys, B., Lust, N. and Granval, P. (1992). Effects of grassland afforestation with different
- 521 tree species on earthworm communities, litter decomposition and nutrient status. Soil

- 522 *Biology and Biochemistry*, 24 (12), p.1459–1466.
- Neilson, R. and Boag, B. (2003). Feeding preferences of some earthworm species common
 to upland pastures in Scotland. *Pedobiologia*, 47 (1), p.1–8.
- 525 Pigott, C. D. (1989). The growth of lime (Tilia cordata) in an experimental plantation and its
- influence on soil development and vegetation. *Quarterly Journal of Forestry*, 83 (1), JOUR,
 p.14–24.
- 528 R Core Team. (2015). *R: A language and environment for statistical computing*.
- 529 Rajapaksha, N. S. S., Butt, K. R., Vanguelova, E. I. and Moffat, A. J. (2013). Earthworm
- 530 selection of Short Rotation Forestry leaf litter assessed through preference testing and
- direct observation. *Soil Biology and Biochemistry*, 67, article, p.12–19.
- Reich, P. B., Oleksyn, J., Modrzynski, J., Mrozinski, P., Hobbie, S. E., Eissenstat, D. M.,
- 533 Chorover, J., Chadwick, O. A., Hale, C. M. and Tjoelker, M. G. (2005). Linking litter calcium, 534 earthworms and soil properties: A common garden test with 14 tree species. *Ecology*
- 535 *Letters*, 8 (8), p.811–818.
- van Rhee, J. A. (1963). Earthworm activities and the breakdown of organic matter in
- 537 agricultural soils. In: Doeksen, J. and van der Drift, J. (eds.), Soil Organisms: Proceedings of
- the Colloquium on Soil Fauna, Soil Microflora and their Relationships. North-Holland,
- 539 Amsterdam: North-Holland Publishing Company, p.55–59.
- van Rhee, J. A. (1969). Development of earthworm populations in polder soils. *Pedobiologia*,
 9, article, p.133–140.
- 542 Rief, A., Knapp, B. A. and Seeber, J. (2012). Palatability of selected Alpine plant litters for the
- 543 decomposer Lumbricus rubellus (Lumbricidae). *PLoS ONE*, 7 (9), p.1–7.
- 544 RStudio Team. (2015). *RStudio: Integrated Development for R*.
- 545 Satchell, J. E. (1983). Earthworm ecology in forest soils. In: Satchell, J. E. (ed.), *Earthworm*
- 546 *Ecology*, London: Chapman & Hall, p.161–170.
- Satchell, J. E. and Lowe, D. G. (1967). Selection of leaf litter by Lumbricus terrestris. *Progress in soil biology*, p.102–119.
- 549 Swift, M. J., Heal, O. W. and Anderson, J. M. (1979a). *Decomposition in terrestrial* 550 *ecosystems*. Oakland, California: University of California Press.
- 551 Swift, M. J., Heal, O. W. and Anderson, J. M. (1979b). *Decomposition in terrestrial* 552 *ecosystems*. Oakland, California: University of California Press.
- 553 Wright, M. A. (1972). Factors governing ingestion by the earthworm Lumbricus terrestris
- (L.), with special reference to apple leaves. *Annals of Applied Biology*, 70 (2), JOUR, p.175– 188.

557 Figure 1.



559 Table I.

Parameter	Soil type		
	Kettering Loam	Reclaimed Soil	
pH (H ₂ O)	7.85 ± 0.03	8.13 ± 0.02**	
Cond. (µs/cm)	748.0 ± 31.3	1558.7 ± 98.0**	
Total N (%)	0.27 ± 0.00	$0.21 \pm 0.00^{***}$	
Total C (%)	3.04 ± 0.02	4.56 ± 0.09***	
C (Org) (%)	2.73 ± 0.03	3.41 ± 0.04***	
O.M. (%)	4.71 ± 0.05	5.88 ± 0.07***	
C (org):N ratio	10.01 ± 0.11	16.06 ± 0.12***	
K (mg/kg)	187.4 ± 1.8	460.8 ± 1.4***	
Са	4324.1 ± 3.3	3933.4 ± 64.6**	
Mg	119.8 ± 0.2	121.0 ± 0.7	
Na	23.55 ± 0.17	19.65 ± 0.51**	
Texture	Clay loam	Sandy clay loam	



562 Figure 2.





b)









570 Table 3.

Soil	Earthworm Species	Days taken to remove 50% total litter	Tree species	
			A. cordata	A. platanoides
Loam	A. longa	15	13.9 ± 5.9	59.9 ± 6.9 ^{***}
	A. chlorotica	Not achieved	92.5 ± 1.3	94.0 ± 0.9
	Mixed Sp.	21	30.3 ± 7.9	41.9 ± 10.5
Reclaimed	A. longa	15	38.3 ± 3.3	$69.7 \pm 9.5^*$
	A. chlorotica	Not achieved	91.1 ± 1.3	93.9 ± 0.9
	Mixed Sp.	21	36.1 ± 7.2	43.9 ± 8.0

571 Student's t-test, n = 5, * p = <0.05, *** p = <0.001.

573 Table 4.

Variable	A. cordata		А. р	A. platanoides	
Variable	Start	End	Start	End	
Total N (%)	2.76	3.62	1.59	2.27	
Total C (%)	52.60	54.90	47.86	48.10	
C:N	19.06	15.15	30.16	21.15	
P (%)	0.13	0.14	0.15	0.17	
Ca (%)	1.16	1.42	1.98	2.38	
К (%)	0.95	0.84	1.21	1.20	
Mg (%)	0.20	0.22	0.22	0.26	

576 Table 5.

Earthworm species	Tree litter preference order
A. longa	ALg, FRe, BEp, ALc > EUn, ACp > ACps, CAs
A. chlorotica	ALg, FRe, BEp > EUn, ACps > ALc, ACp , CAs

579 Table and figure captions

Table 1. Mean selected parameters (± SE) of reclaimed soil and Kettering loam, prior to use in the earthworm choice chamber experiment. ANOVA, n = 3, * p = <0.05, ** p = <0.01, *** p = <0.001.

Table 2. Initial and final (after 27 days) mean parameters of monocultures and mixed
cultures of the earthworms *Aporrectodea longa* and *Allolobophora chlorotica* in choice
chambers containing reclaimed soil or Kettering loam.

- Table 3. Mean (± SE) remaining *Alnus cordata* and *Acer platanoides* foliar material (% from original mass, wet basis) in choice chambers containing monocultures or mixed cultures of the earthworms *Aporrectodea longa* and *Allolobophora chlorotica* and reclaimed soil or Kettering loam, at the point of 50% total foliar material removal. ANOVA, n = 5, * p = <0.05, *** p = <0.001.
- Table 4. Chemical analysis of *Alnus cordata* and *Acer platanoides* foliar material at the start and termination of the earthworm choice chamber experiment (after 27 days), n=1.
- 593 Table 5. Tree litter and foliar preference by the earthworms *Aporrectodea longa* and
- 594 Allolobophora chlorotica following Rajapaksha et al. (2013), updated with the results of this
- 595 choice chamber experiment (in bold) as appropriate for Kettering Loam. Tree species: *Alnus*
- 596 glutinosa (ALg), Fraxinus excelsior (FRe), Betula pendula (BEp), Eucalyptus nitens (EUn),
- 597 Castanea sativa (CAs), Acer pseudoplatanus (ACps), Alnus cordata (ALc) and Acer
- 598 platanoides (ACp).

599 Figure 1. a) Empty choice chamber prior to use in an earthworm-based foliar preference

- 600 experiment, b) detail of empty Eppendorf tube food vessel fixed to the wall of a choice 601 chamber via drilled cap.
- Figure 2. Prepared earthworm choice chambers, each containing *Alnus cordata* and *Acer platanoides* foliar material and a soil treatment: a) Kettering loam, b) reclaimed soil (with
- 604 individuals of the earthworm species Allolobophora chlorotica on soil surface immediately
- 605 after addition).
- Figure 3. Mean (± SE) foliar mass remaining (% wet basis) in choice chambers over a period
- of 27 days. Earthworm species combinations as labelled, in (a) loam soil and (b) reclaimed
- soil. Tree foliar species: *Acer platanoides* (●) and *Alnus cordata* (■).